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The effect of warming climate on nutrient and solute export from the Greenland Ice Sheet

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Abstract

Glacial meltwater runoff is likely an important source of limiting nutrients for downstream primary producers. This has particular significance for regions surrounding the Greenland Ice Sheet, which discharges >400 km³ of meltwater annually. The Arctic is warming rapidly but the impact of higher discharge on nutrient export is unknown. We present four years of hydrological and geochemical data from a large Greenland Ice Sheet catchment that includes the two highest melt years on record (2010, 2012). Measurements reveal significant variation in dissolved solute (major ion) and estimated dissolved macronutrient (nitrogen, phosphorus and silica) fluxes, with increases in higher melt years. Labile particulate macronutrients dominate nutrient export, accounting for ~50 % of nitrogen and >80 % of both phosphorus and silica. The response of ice sheet nutrient export to enhanced melting is largely controlled by particle bound nutrients, the future supply of which is uncertain. We propose that the Greenland Ice Sheet provides an underappreciated and annually dynamic source of nutrients for the polar oceans, with changes in meltwater discharge likely to impact marine primary productivity in future decades.

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Introduction

Recent estimates predict global mean surface warming of up to 4.8 °C above the 1986–2005 mean by 2100, with the polar regions subject to more extreme increases (Collins *et al.*, 2013). Already, the Greenland Ice Sheet has experienced

increased surface temperatures, with the five highest melt seasons on record occurring since 2000 (Tedesco *et al.*, 2013). In 2012, surface melting was the most widespread in over 100 years (Tedesco *et al.*, 2013). By 2100, the annual freshwater flux from the Greenland Ice Sheet could exceed 1000 km³ a⁻¹, making it one of the world's largest sources of freshwater (Fettweis *et al.*, 2013).

Currently, we lack information about the impact of meltwater on downstream biogeochemical cycles, even though the coastal waters surrounding the ice sheet harbour highly productive ecosystems, that are strong CO₂ sinks (Rysgaard *et al.*, 2012). Recent work has highlighted the importance of glacier meltwater, including delivery of essential nutrients to the polar oceans (Bhatia *et al.*, 2013; Wadham *et al.*, 2013; Hawkings *et al.*, 2014; Lawson *et al.*, 2014). However, whether glacier melting provides an important negative climate feedback through its effect on marine primary production and CO₂ drawdown remains unknown.

Future changes to Greenland Ice Sheet hydrology will probably impact the export of solute and reactive sediments to the polar oceans. Much of the meltwater drains from the surface to the ice sheet bed, chemically weathering the subglacial sediments (Bartholomew *et al.*, 2011). Supraglacial lake drainage events are particularly important because they rapidly channel large quantities of meltwater to the ice sheet bed, flushing out stored, solute-rich, subglacial waters (Bartholomew *et al.*, 2011; Hawkings *et al.*, 2014). Supraglacial lake formation and the migration of drainage systems inland in a warming climate (Leeson *et al.*, 2015) could expose new subglacial areas to meltwater flushing, potentially enhancing solute evacuation. Glaciers are effective at fracturing and grinding bedrock (Cowton *et al.*, 2012), producing turbid meltwaters with abundant, very fine suspended particles, *i.e.* >1 g L⁻¹ (Cowton *et al.*, 2012). Suspended material has recently been identified as a potential source of labile nutrients to near coastal regions (Hodson *et al.*, 2004; Bhatia *et al.*, 2013; Hawkings *et al.*, 2014; Wehrmann *et al.*, 2014) but data are sparse.

In this study, we present a full suite of geochemical and hydrological data from Leverett Glacier, a large (~600 km²), land terminating, outlet glacier of the Greenland Ice Sheet (details in Supplementary Information). The data cover four years (2009–2012) where melting intensity varied (Fig. 1), including the two highest melt seasons on record. This is the most comprehensive dataset yet on major ion and nutrient concentrations from a glacial system.

Results and Discussion

Hydrological data (discharge, electrical conductivity and suspended material concentration) were collected for all four years (2009–2012) at a stable bedrock section ~2.2 km downstream of the glacier mouth. Major ion data are available for 2009, 2010 and 2012 and we used them to interpolate concentrations for 2011. Nutrient flux for 2009, 2010 and 2011 was estimated from 2012 data using a correlation with electrical conductivity. Further details are provided below and in Supplementary Information.

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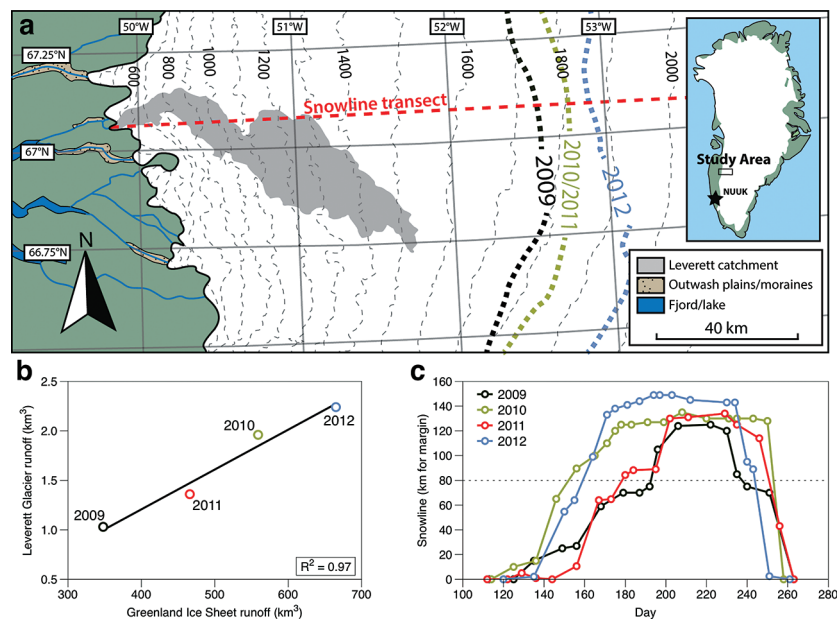


Figure 1 Field site and basic hydrological data. (a) Leverett Glacier with the snowline transect used to determine the point where snow covers the ice surface, derived from MODIS imagery. Dashed lines represent elevation contours. The catchment area was determined from a surface digital elevation model (Palmer *et al.*, 2011). (b) Leverett Glacier meltwater discharge (Table 1) versus modelled Greenland Ice Sheet runoff (Tedesco *et al.*, 2013). (c) Leverett Glacier snowline; open circles represent observed position; connecting lines are linear estimates of retreat, interpolated between the measurements. Estimated catchment extent is represented by the dotted line at ~80 km.

Hydrology

The 2010 and 2012 ablation seasons produced the largest volumes of meltwater on record (Tedesco *et al.*, 2013). This is reflected in Leverett Glacier discharge (Fig. S-1), which was proportional to annual ice sheet runoff ($R^2 = 0.97$; Fig. 1b). The snowline also reached maximum elevation in 2012, 14 km further inland than in 2010 and 2011 (Table 1; Fig. 1c). 2009 and 2011 can be considered “average” melt years, with discharge proportional to the mean meltwater flux over the past decade. 2010 and 2012, with significantly above average discharge, were “extreme” melt years. This characterisation provides a benchmark for evaluating future trends because the frequency of extreme seasons is likely to increase (Fettweis *et al.*, 2013).

Table 1 Flux and hydrological data.

	Units	Year			
		2009	2010	2011	2012
Greenland Ice Sheet runoff *	km ³	348	559	466	665
Leverett Glacier discharge	km ³	0.94	1.79	1.10	2.03
Snowline above catchment	est. days	45	100	73	85
Snowline retreat from margin	km	125	135	135	149
Solute Flux	eq	3.0×10^8	5.6×10^8	3.2×10^8	5.6×10^8
Sediment Flux	t	3.7×10^6	2.6×10^6	3.0×10^6	2.2×10^6
Dissolved inorganic nitrogen †	t	25	46	26	48
Dissolved silica †	t	130	230	120	230
Dissolved inorganic phosphorus†	t	6.2	12	6.7	12
Dissolved inorganic nitrogen ‡	t	22	41	25	47
Dissolved silica ‡	t	110	220	130	240
Dissolved inorganic phosphorus‡	t	7.8	15	8.9	16
Exchangeable NH ₄ **	t	19-58	13-41	15-47	11-35
Amorphous silica**	t	18,000-44,000	12,000-31,000	14,000-36,000	11,000-26,000
NaOH extractable phosphorus**	t	20-130	14-92	16-110	12-78

All estimates are shown to Decimal Day 230/231, i.e. 17 August.

Estimates are reported with 2 significant digits.

eq = molar equivalent

t = tons of dry element

Snowline: the boundary where snow covers the underlying ice. Down glacier from this point is exposed ice, where the snow cover has melted.

* Greenland Ice Sheet runoff estimates from Tedesco *et al.* (2013).

** Sediment fluxes given as range based on minimum and maximum extractable nutrient concentrations.

† Fluxes estimated with electrical conductivity.

‡ Fluxes estimated using discharge weighted mean.

Solute flux

We estimated total solute export for all years from the electrical conductivity (EC) of the meltwater (Fig. S-3). The major ion (Ca^{2+} , Mg^{2+} , K^+ , Na^+ , Cl^- , SO_4^{2-} , HCO_3^-) concentrations are a linear function of conductivity with $R^2 = 0.87$ (2009), 0.77 (2010) and 0.98 (2012) (Figs. S-3 and S-4). We have a full hydrological dataset (discharge, EC and suspended material) from 2011 but major ion data are missing so using the concentration-conductivity correlation for data from 2009, 2010 and 2012 ($n = 368$; $R^2 = 0.89$), we estimated 2011 solute concentrations (Fig. S-3). Solute flux was estimated by cumulatively summing the total solute concentration



at each conductivity measurement time step (Fig. 2). We differentiated supraglacial solute from the total solute flux to assess the importance of subglacial sources (Fig. 2a). The estimates suggest that flushing of stored subglacial waters and rapid weathering of subglacial sediments by dilute supraglacial meltwater account for >95 % of the solute export from the catchment.

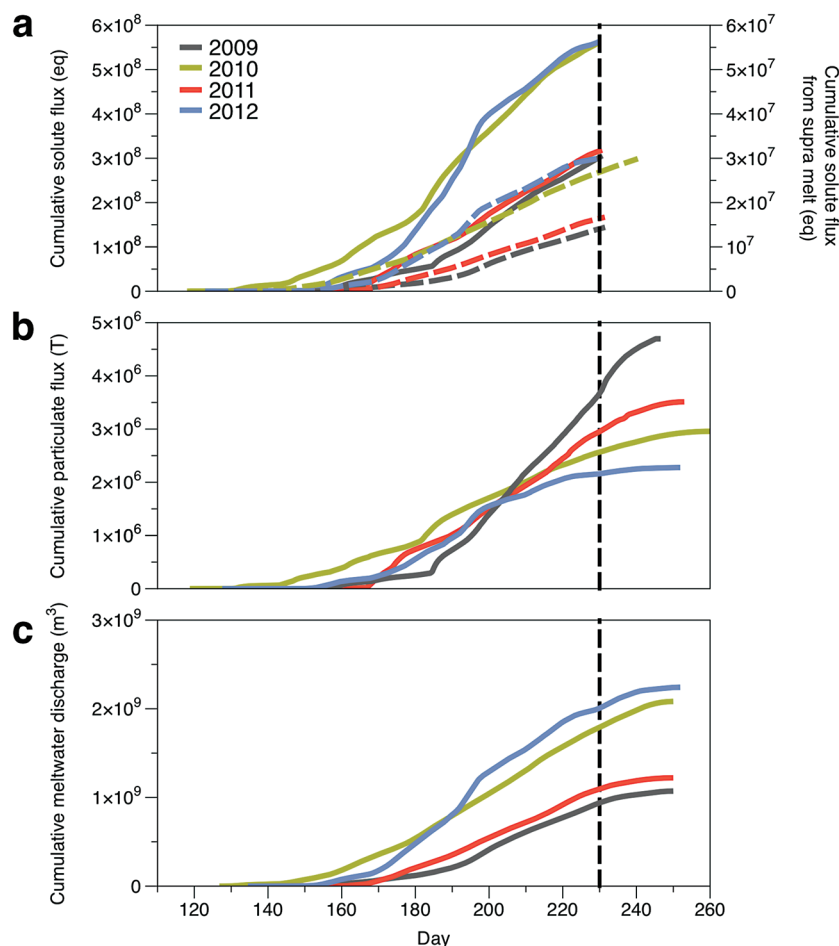


Figure 2 Cumulative solute and sediment flux. (a) Solid lines show solute flux (Ca^{2+} , Mg^{2+} , K^+ , Na^+ , Cl^- , SO_4^{2-} , HCO_3^-); dashed lines represent the estimated portion originating from supraglacial meltwater (right axis). (b) Cumulative particle borne flux. (c) Cumulative meltwater discharge. Day 230 (dashed line) was the limit of the dataset interpretation. Values at the dashed line correspond to data presented in Table 1 and Figure 1b.

A key discovery is that solute flux during the two extreme melt years was ~90 % higher than for the average years, indicating that solute export scales with discharge. Thus, it is likely that increased melting will increase solute fluxes from the Greenland Ice Sheet. Most of the solute is delivered during peak melt periods during the ablation season (Fig. S-5). Solute pulses are also released during supraglacial lake drainage events (Fig. S-5; Bartholomew *et al.*, 2011). Enhanced solute discharge during extreme years likely results from i) drainage of concentrated subglacial waters stored in poorly connected regions, *e.g.*, deeper into the ice sheet, ii) higher flushing rates and iii) rapid weathering of reactive subglacial sediments by large volumes of supraglacial meltwater. We next ascertain if nutrient fluxes follow the same trend.

Dissolved nutrient fluxes

Nutrient fluxes for all years were estimated from the 2012 dataset, the only complete macronutrient record available. For the most common, namely NH_4^+ , NO_3^- , Si and PO_4^{3-} , we also estimated annual flux by correlation with conductivity, as for major ion concentrations. This is justified by the good correlation of conductivity with nutrient concentrations (Fig. S-6). If the function holds for 2012, we assume it also holds for the other years. Si and P are released during rock weathering and dissolved inorganic nitrogen is also enhanced by subglacial biogeochemical processes, such as microbially mediated nitrification (NO_3^-) and mineralisation of organic matter (NH_4^+ ; Wynn *et al.*, 2007). For example, mean nitrogen concentrations attributable to subglacial processes (~1.3 μM) are similar to supraglacial processes (~1.3 μM). Nitrogen correlates linearly with conductivity ($R^2 = 0.74$; Fig. S-6). To account for the seasonal evolution in meltwater composition, we used two regressions (Fig. S-6) for dissolved silica ($R^2 = 0.72$ for early season and $R^2 = 0.34$ for bulk season runoff), and phosphate ($R^2 = 0.61$ and 0.59, respectively). We propose that these differences arise from the change in source of the subglacial water as the melt season progresses, *i.e.* close to ice margins early in the season to more isolated inland subglacial waters as the season progresses. The source influences the subglacial flowpath length, hence water residence times and pH/redox conditions.

Higher dissolved nutrient flux correlates with higher discharge years. Inorganic nitrogen (86 % \pm 9.8 %), dissolved silica (85 % \pm 0.1 %) and phosphate (86 % \pm 11.9 %) are higher in extreme melt years than for average years. This is significant and demonstrates the potential for nutrient release by a warming climate.

A different approach for estimating nutrient flux is to combine discharge weighted mean concentrations and total discharge flux (Hawkins *et al.*, 2014). Estimates from this method and the EC-based method are similar (Table 1). The weighted mean dissolved nitrogen and silica fluxes were marginally lower (~7 % and ~5 %) and phosphate flux was higher (~23 %), probably because of late melt season influence, when phosphate concentrations were high and where the bulk of the discharge occurred.



An important assumption is that the correlation of nutrient concentration with conductivity is consistent over the years. We have a limited dataset for NO_3^- and Si from 2009. Results are sparse so estimates are crude but they serve as a benchmark for comparison. Flux derived from 2009 data for NO_3^- N is 12 tons, compared with 15 tons estimated using 2012 data. Flux for dissolved Si is 180 tons, compared with 130 tons estimated using 2012 measurements (Table 1). Both are well within an order of magnitude, which offers confidence that our estimates from 2012 data are reasonable.

Nutrient flux on particles

Glaciers effectively fracture and grind bedrock into high surface area, highly reactive, clay and silt sized particles, some of which are transported in runoff as suspended material (Gurnell and Clark, 1987; Brown *et al.*, 1996). By using data derived from the labile nutrients in the 2012 suspended material ($n > 25$), we estimated the range of nutrient concentrations associated with the particulate fraction (Table 1; Fig. 3). We assumed that the 2009–2011 mean extractable nutrient concentrations lie within the 2012 minimum and maximum concentrations, which is reasonable because the runoff comes from the same catchment and the mineral composition is relatively constant (Hawkins *et al.*, 2014).

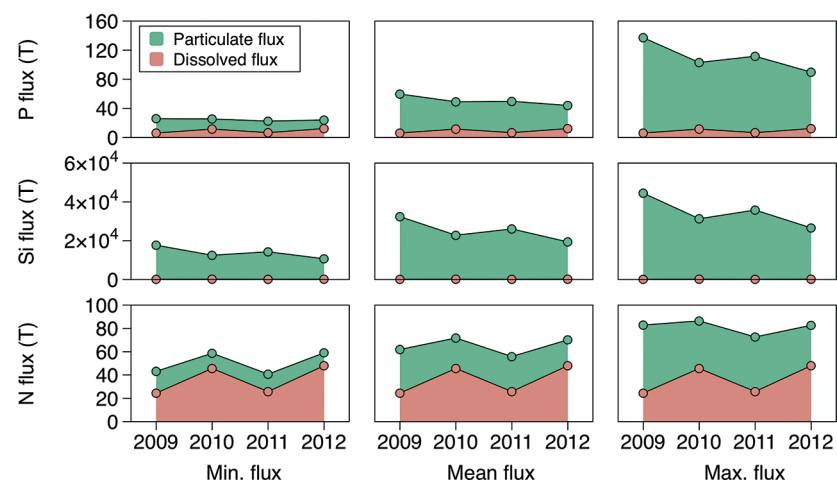


Figure 3 Estimated nutrient flux. Minimum (left column), mean (middle) and maximum (right) possible values give an impression of the range for phosphate, silicate and nitrogen compounds. Dissolved flux was determined using the electrical conductivity correlation.

Particulate bound nutrients account for a large portion of the estimated nutrient flux (Table 1; Fig. 3), which is significant in all years. The particulate transported fraction correlates with the nutrient source (lithogenic or atmospheric)

and the tendency for the ion to complex. Nutrients derived directly from rocks associate more with solids, *i.e.* for Si, >99 % and for P, ~80 % of the total flux. Nitrogen, which has a supraglacial component, is transported less on solids, *i.e.* <50 %. Our results are consistent with the low solubility of Si and the high affinity for P absorption onto solids, such as iron (oxyhydr)oxides. Ammonium is only weakly absorbed and nitrate remains preferentially in solution. This suggests that annual nitrogen flux is more sensitive to changes in ice sheet water discharge than particle flux.

We have demonstrated that, as in riverine systems (Mayer *et al.*, 1998; Ruttenberg, 2014), a high fraction of ice sheet nutrient export is associated with suspended material. This is consistent with previous research from Arctic glaciers (Hodson *et al.*, 2004) and supports recent assertions that the impact of terrigenous material on the oceans is underestimated in global element cycling (Jeandel and Oelkers, 2015). Our results underline the need for more information about ice sheet sediment flux dynamics. As in previous studies, we observed highly variable annual sediment fluxes, which do not correlate well with discharge on a catchment basis (Fig. 2; Gurnell and Clark, 1987). Sediment flux might be less influenced by total meltwater discharge and more sensitive to meltwater access to fresh, subglacial sediment sources (Cowton *et al.*, 2012). However, evidence from past deglaciation events indicates that climate warming increases sediment export (Jeandel and Oelkers, 2015) and analysis of sediment plumes from meltwater rivers demonstrates a higher sediment flux in recent years (Hudson *et al.*, 2014).

The extent of biological consumption of the nutrients bound to particles before deposition and subsequent burial is unknown. Particulates from meltwater are extremely fine, *e.g.*, >95 % of particles can be <32 μm in size (Brown *et al.*, 1996), so surface area is high and nutrient transport in the buoyant, fresh water plumes in near coastal regions is likely to be significant. Evidence from recent polar studies shows that particle borne nutrients are carried far offshore (Schroth *et al.*, 2014; Wehrmann *et al.*, 2014) and nutrients deposited with glacial sediments in fjords can be resuspended in the water column (Wehrmann *et al.*, 2014). However, the scarcity of data means that the contribution of particle bound nutrients on oceanic productivity near Greenland remains uncertain.

Terrestrial and marine studies have shown that large fractions (75–95 %) of amorphous Si can be dissolved and recycled (Treguer *et al.*, 1995; Gibson *et al.*, 2000). Amorphous Si is an order of magnitude more soluble in saline solutions than in fresh water (Icenhauer and Dove, 2000; Loucaides *et al.*, 2008) and recycling is favoured in estuaries (Loucaides *et al.*, 2008). NaOH extractable phosphate is commonly termed “algae available” (DePinto *et al.*, 1981) and its bioavailability has previously been demonstrated (Bostrom *et al.*, 1988). High salinity in ocean and fjord waters also favours P and NH_4 desorption (Garner *et al.*, 1991; Hodson *et al.*, 2004; Zhang and Huang, 2011), enhancing their bioavailability. Thus, annual sediment flux is an important factor in downstream productivity.



Conclusions

Changes in the hydrological output from the Greenland Ice Sheet in a warming climate could have significant effect on solute and nutrient delivery to near coastal regions. Our data, from Leverett Glacier, a large representative ice sheet catchment, indicate that bulk solute and dissolved nutrient fluxes will increase as “extreme” melt year frequency increases. A significant fraction of nutrients, especially silica and phosphorous, will be transported by suspended particles. The extent of their influence depends on desorption before burial, bioavailability and change in the ice sheet particulate flux, which are currently uncertain. Our study demonstrates that retreating snowline and higher meltwater input into less efficiently drained subglacial regions are likely to increase the dissolved macronutrient flux. Particle bound nutrients have been largely overlooked but contribute significant mass to nutrient cycling. Increased warming, thus increased meltwater runoff, will likely impact regional nutrient availability, and thus, the carbon cycle.

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Additional Information

Supplementary Information accompanies this letter at www.geochemicalperspectivesletters.org/article1510

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Author Contributions

All authors contributed. JLW and MT conceived the project. JRH, EL, AS, TC, IB, AT, PN, DC and JLW collected field data. JRH, JRH, EL and JT undertook the lab analysis. JRH, JLW and MT wrote the paper.

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■ Supplementary Information

The Supplementary Information includes:

- Study Site and Methods
- Figures S-1 to S-6
- Supplementary Information References

Study Site and Methods

Study area

Research was conducted at Leverett Glacier (LG; Fig. 1a; 67.06 °N, 50.17 °W), a large, land terminating glacier on the southwestern margin of the Greenland Ice Sheet (GrIS). The catchment extends >80 km into the ice sheet and is estimated to cover an area of >600 km² (Palmer *et al.*, 2011; Cowton *et al.*, 2012). Leverett overlies bedrock of Archean gneiss and granite, common to much of Greenland (Henriksen *et al.*, 2009). Catchment hydrology is typical of large Greenland outlet glaciers and is described elsewhere (Chandler *et al.*, 2013). A catchment hydrological record was maintained during the 2009–2012 summer ablation seasons, with monitoring of discharge, electrical conductivity (EC) and turbidity (suspended material), recorded at 5–10 minute intervals (Fig. S-1).

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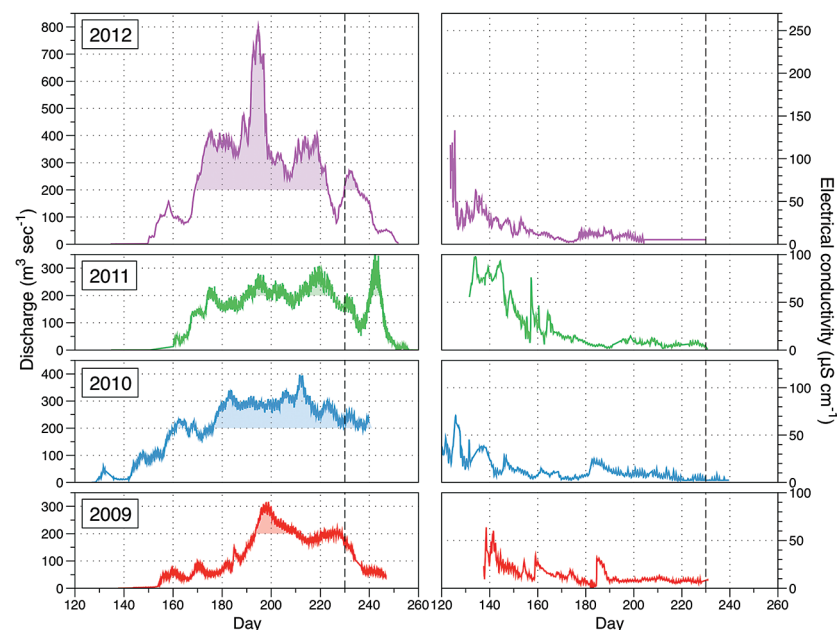


Figure S-1 Leverett Glacier discharge hydrographs and electrical conductivity. Shaded areas in the discharge plots show when discharge exceeded $200 \text{ m}^3 \text{ sec}^{-1}$. The dashed line represents the cut off date used for flux estimates.

Water sample collection and filtration

Bulk meltwater samples for geochemical analysis were collected at least once a day (2009, 2010, 2012), from a sampling site located ~1 km (2012) or ~2 km (2009, 2010) downstream from the Leverett Glacier terminus, throughout the main melt period (May–August). Our confidence that these waters represent the bulk discharge is based on the drainage of Leverett from a single portal. The composition of point samples taken there were within the uncertainty of those taken further downstream.

Grab samples were immediately passed through 47 mm $0.45 \mu\text{m}$ cellulose nitrate filters (Whatman®), mounted on a PES filtration stack (Nalgene™), that had been rinsed 3 times with the sample. The filtrate was immediately frozen in clean HDPE 30 mL Nalgene™ bottles that had been rinsed 3 times with the filtered sample. Major ion analysis (Ca^{2+} , Mg^{2+} , K^+ , Na^+ , Cl^- , SO_4^{2-} , HCO_3^-) and NO_3^- was completed within three months of collection. We used a Thermo Scientific™ Dionex™ DX-500 (2009, 2010) or a Thermo Scientific™ Dionex™ capillary ICS-5000 (2012), fitted with simultaneous anion and cation columns.

Measurement accuracy was $\pm 4\%$ and precision was $\pm 7\%$ for the DX-500, and $\pm 3\%$ and $\pm 3\%$ for the ICS-5000. Dissolved macronutrients were measured from the 2012 samples using a LaChat QuickChem® 8500 flow injection analyser system using low level detection methods (Si , PO_4^{3-}) or by manual colorimetric techniques (NH_4^+), as described by Le and Boyd (2012). All samples were field blank corrected where the blank concentrations were above the detection limit of the instrument. In total, 408 samples were analysed for major ions and 75 samples, for nutrient concentrations. In 2012, snow and supraglacial meltwaters draining into a moulin ~30 km from the catchment margin were sampled for geochemical analysis ($n = 32$), using the same methods as the bulk geochemical samples. Snow samples were placed in new Whirl-Pak® bags (Nasco) and left in a water bath to melt. Samples were filtered, as above, as soon as melting was complete.

Sediment nutrient extractions

Particulate bound nutrient samples were taken during the 2012 melt season ($n = 25$ for P and Si, $n = 39$ for N). Briefly, a meltwater sample of 300–400 mL was filtered through a $0.45 \mu\text{m}$ cellulose nitrate filter (PSi and PP; Whatman®) or a $0.7 \mu\text{m}$ glass microfibre filter. Suspended particulate material was retained for the commonly used labile nutrient extractions, “algae available” P (Hodson *et al.*, 2004), exchangeable NH_4 (PN; Maynard *et al.*, 2007), and amorphous Si (PSi; DeMaster, 1981). Particulate material was removed carefully by gentle scraping from the filter, and weighed. Mean extractable concentrations were combined with the total sediment flux from Leverett Glacier, to determine the labile particulate nutrient flux.

“Algae available” phosphorus extraction

Owing to its importance as an essential nutrient, phosphorus extraction techniques are well documented in the literature (Dorich *et al.*, 1980; DePinto *et al.*, 1981; Sharpley *et al.*, 1991; Ekholm and Krogerus, 2003; Hodson *et al.*, 2004). Here we used a common extraction method that aims to determine the amount that is bioavailable (Bostrom *et al.*, 1988; Hodson *et al.*, 2004). We adapted the standard method to allow for analysis of very small quantities of sediment, *i.e.* “micro-extraction”. Briefly, 1.5 mL of 0.1 M NaOH solution was added to ~50 mg of sediment, that was accurately measured to $\pm 0.001 \text{ g}$ using a high precision/accuracy balance. We used 2 mL microcentrifuge tubes, which allow similar sediment to extractant ratios to those used by others (Hodson *et al.*, 2004). Microcentrifuge tubes were capped and placed on a reciprocating (rotary) shaker at 200 rpm for 16 hours. Tubes were then centrifuged at 2600 rpm for 10 minutes, the supernatant was transferred to new 2 mL tubes using a 1 mL plastic syringe (PP/PE) and filtered through a syringe filter (Whatman® Puradisc PP).



Exchangeable ammonium extraction

We adopted the method described by Maynard *et al.* (2007) and applied elsewhere (Telling *et al.*, 2011, 2012, 2014). 10 mL of 2.0 M KCl solution was added to 0.7 µm filters (GF/F Whatman®) in a 15 mL plastic centrifuge tube. Tubes were capped and placed on a reciprocating (rotary) shaker (160 rpm) for 30 minutes. Solutions were then decanted into fresh plastic tubes, centrifuged, filtered through a 0.45 µm syringe filter (Whatman® Puradisc PP) and frozen at -20 °C until analysis. All sediment from filters was retained, rinsed with Milli-Q deionised water (18.2 MΩ cm⁻¹ Millipore) to remove extract solution and dried in an oven overnight at ~50 °C to provide dry sediment weights. These were cross checked against weights expected from hydrological suspended sediment records, *i.e.* 300 mL of water was filtered so an expected weight could be generated from the recorded meltwater suspended material. Nine blanks were treated in the same manner as the samples, using the same types of filter, to test for filter contamination.

Amorphous silica extraction

This is the first study to present data on extraction of amorphous silica in glacial sediments. Here we use a method commonly employed in marine and riverine systems that was developed by DeMaster (1981) and validated for terrestrial soils and sediments (Sauer *et al.*, 2006). The technique uses 0.1 M Na₂CO₃, a weak base, which maximises dissolution of amorphous Si, with minimal impact on crystalline material. About 30 mg of sediment was accurately weighed into a 60 mL HDPE bottle (Nalgene®) and 50 mL of 0.1 M Na₂CO₃ solution was added. Bottles were placed in a hot water bath (85 °C) and 1 mL aliquots were removed after 1, 2, 3 and 5 hours. Aliquots were refrigerated in 2 mL microcentrifuge tubes at 4 °C until analysis, less than 24 hours later. Just prior to analysis, 0.5 mL of sample was neutralised with 4.5 mL 0.021 M HCl in plastic centrifuge tubes. Three blanks were processed alongside the samples to check for method contamination. Amorphous silica was determined by using the intercept of the regression line drawn through Si concentrations obtained from the time series aliquots; amorphous silica dissolves within the first hour of the extraction procedure (DeMaster, 1981).

Analysis of extract solutions and filtered meltwater

Phosphorus and silica extraction solutions were measured on a LaChat Quick-Chem® 8500 flow injection analyser system (Method Numbers 31-115-01-1-I for P and 31-114-27-1D for Si). The coefficient of variation (CoV) for the method was ±0.5 % for silica (based on seven replicate standards) and ±3.2 % for dissolved orthophosphate (five replicate standards). Limits of detection were 0.3 µM (8.4 µg Si L⁻¹) and 0.01 µM (0.3 µg P L⁻¹). Ammonium was determined using a Bran and Luebbe Autoanalyzer 3 (for extractants) or by the manual salicylate method (Le and Boyd, 2012). The Luebbe Autoanalyzer 3 method had a CoV of ±1.5 % (eight

replicate standards) and a detection limit of 0.5 µM (6.4 µg N L⁻¹). The manual method had a CoV of ±4.9 % (five replicate standards) and a limit of detection of 0.6 µM (8.4 µg N L⁻¹). All samples were blank corrected where blank concentrations were higher than the detection limits.

Contribution of supraglacial solute to total solute flux

Measured mean major ion concentration in supraglacial melt (snow and ice melt from 2012, n = 32) was 15 µeq L⁻¹ (±7 µeq L⁻¹). There is no reason to expect there to be significant annual variation in these concentrations, which are more than an order of magnitude lower than mean bulk meltwaters (~400 µeq L⁻¹). The mean supraglacial major ion value of 15 µeq L⁻¹ was multiplied by catchment discharge at each time step and cumulatively summed (Fig. 2) to derive solute fluxes from supraglacial sources.

Catchment hydrological monitoring and meltwater/sediment fluxes

Leverett Glacier runoff was hydrologically gauged throughout the 2009-2012 summer melt seasons, from late April or early May at the onset of melting, through to late August or early September. Briefly, the discharge, electrical conductivity and turbidity (suspended material concentration) of the meltwater river were logged every 5 minutes at a stable bedrock section ~2.2 km downstream from the glacier terminus (Cowton *et al.*, 2012). Discharge was determined using the method described by Bartholomew *et al.* (2011). A wired water pressure sensor monitored stage, which was converted into discharge using a (stage-discharge) rating curve of Rhodamine WT dye dilution injections. Twenty nine dye dilutions were used in 2009 and 2010, 26 in 2011 and 41 in 2012. Leverett meltwater and particulate material fluxes were determined by multiplying discharge and suspended material concentrations at the 5 minute time points by 300 (to derive values for each second over the 5 minute period between new recordings) and summed over the entire melt season. For comparison, modelled meltwater runoff data for the Greenland Ice Sheet were taken from Tedesco *et al.* (2013).

Snowline determination

Snowline retreat throughout the 2009-2012 melt seasons was monitored using MODIS (moderate resolution imaging spectroradiometer) on the Terra platform. Surface reflectance band 1-2 images (Product Number MOD09GQ) were processed using the QGIS analysis package, to determine snowline extent (Fig. S-2). Maximum snowline position from monitored years is shown in Figure 1. A minimum of 18 cloud free images were used per year to determine snowline migration.



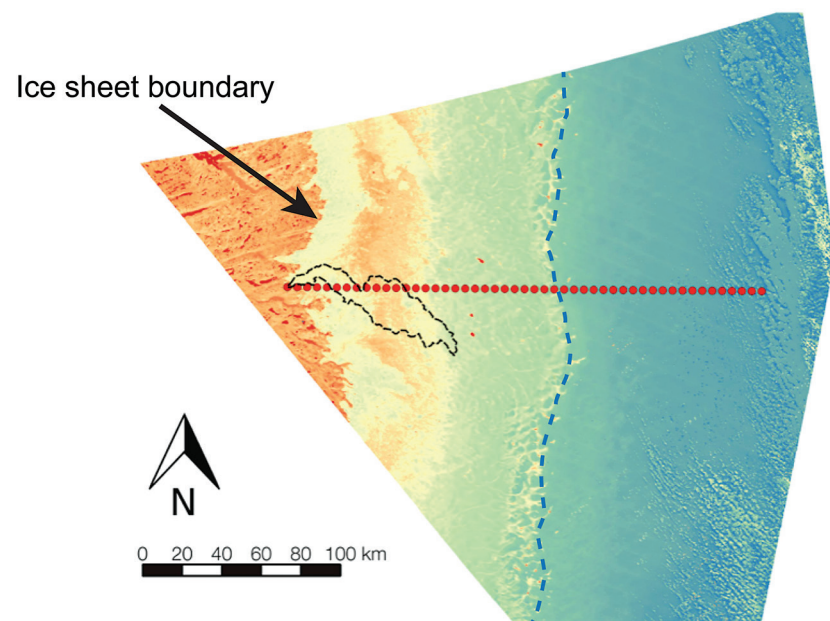


Figure S-2 An example of the MODIS satellite imagery used to determine snowline extent. Data are from surface reflectance bands 1-2 (Product Number MOD09GQ). Leverett Glacier moves from east to west. The estimated catchment area is outlined in black (Palmer *et al.*, 2011) and the snowline transect is displayed by the red dotted line (space between each dot = 5 km). False colour imaging differentiates between snow (dark blue) and ice (light blue). The margin of the ice sheet is evident at the border of the red colouring (black arrow). The blue dashed line represents the interpreted position of the snowline. The image displayed is from 21 July 2012. Fig. 1a provides more information about the catchment.

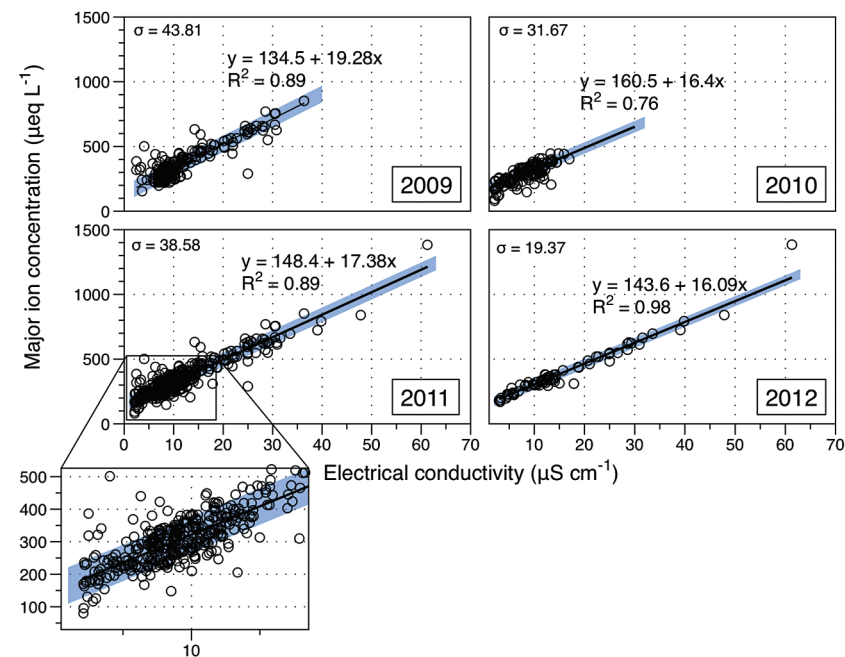


Figure S-3 Regression plots for major ions as a function of electrical conductivity. Shaded blue lines represent the standard error (σ , also written in the top left of plots). The 2011 regression plot (bottom left), which encompasses data from 2009, 2010 and 2012, is magnified to demonstrate the linear relationship of electrical conductivity with major ions in the dilute waters.



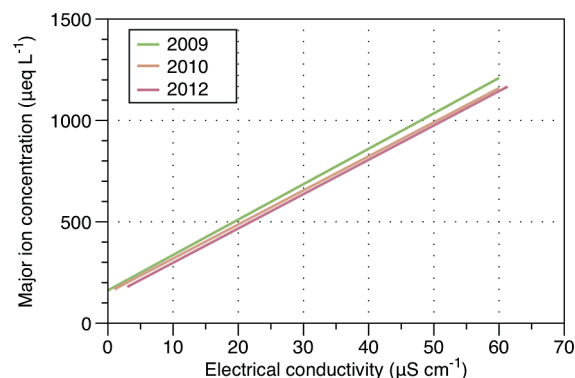


Figure S-4 Comparison of regression plots for the major ions for 2009, 2010 and 2012, where data are available. The differences in the regression equations used to determine major ion concentrations are very small.

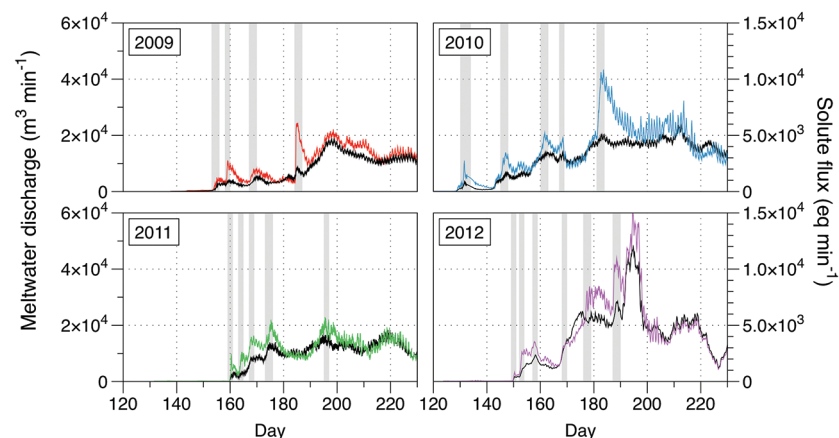


Figure S-5 Temporal variation in meltwater discharge and estimated solute flux. Coloured lines indicate solute flux calculated by the EC based method (eq min^{-1}), and black lines, meltwater discharge ($\text{m}^3 \text{min}^{-1}$), during the 2009–2012 melt seasons. Shaded areas correspond to meltwater pulse events. These are associated with spring events, *i.e.* the annual opening of the subglacial drainage system, or rapid drainage of meltwater from supraglacial lake drainage events. These meltwater pulses flush concentrated waters from the subglacial drainage system (Bartholomew *et al.*, 2011 and Hawkins *et al.*, 2014).

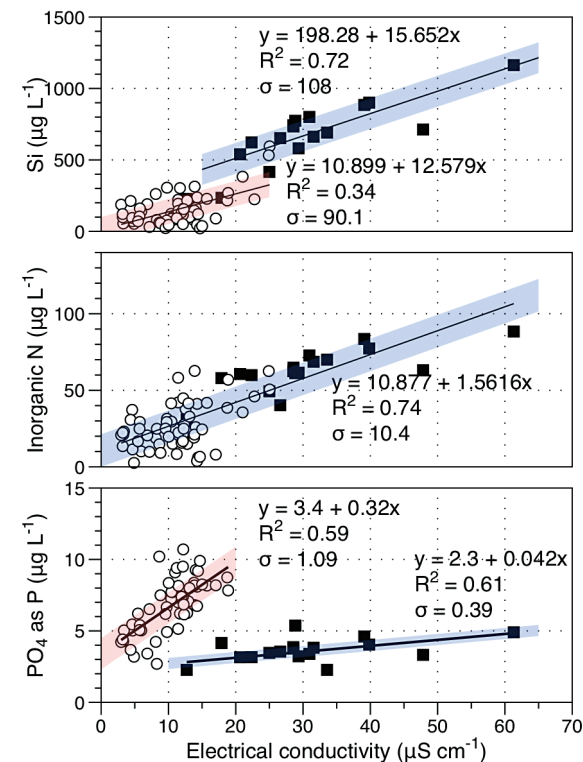


Figure S-6 Regression plots for nutrient concentrations as a function of electrical conductivity. The regression plots for Si and PO_4^{3-} represent early (blue) versus later season (red) hydrological and associated biogeochemical changes. Early season regressions are applied to time points before Day 153 in 2009, Day 128 in 2010, Day 160 in 2011 and Day 150 in 2012, which are before the initial change in drainage hydrology occurred, *i.e.* the spring event (Figs S-1 and S-5). Shaded areas show associated standard error.

Supplementary Information References

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